



Current State of Research on the Risk of Morbidity and Mortality Associated with Air Pollution in Korea

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Purpose: The effects of air pollution on health can vary regionally. Our goal was to comprehensively review previous epidemiological studies on air pollution and health conducted in Korea to identify future areas of potential study.

Materials and Methods: We systematically searched all published epidemiologic studies examining the association between air pollution and occurrence of death, diseases, or symptoms in Korea. After classifying health outcomes into mortality, morbidity, and health impact, we summarized the relationship between individual air pollutants and health outcomes.

Results: We analyzed a total of 27 studies that provided 104 estimates of the quantitative association between risk of mortality and exposure to air pollutants, including particulate matter with aerodynamic diameter less than 10 μ m, particulate matter with aerodynamic diameter less than 2.5 μ m, sulfur dioxide, nitrogen dioxide, ozone, and carbon monoxide in Korea between January 1999 and July 2018. Regarding the association with morbidity, there were 38 studies, with 98 estimates, conducted during the same period. Most studies examined the short-term effects of air pollution using a time series or case-crossover study design; only three co-hort studies that examined long-term effects were found. There were four health impact studies that calculated the attributable number of deaths or disability-adjusted life years due to air pollution.

Conclusion: There have been many epidemiologic studies in Korea regarding air pollution and health. However, the present review shows that additional studies, especially cohort and experimental studies, are needed to provide more robust and accurate evidence that can be used to promote evidence-based policymaking.

Key Words: Air pollution, mortality, morbidity, environmental medicine, Korea

INTRODUCTION

The effect of air pollution on mortality and the burden of disease increases as air pollution increases, although estimates can vary from region to region. According to the Global Burden of Disease Study, ambient air pollution accounted for 7.5% of deaths globally in 2016 and was the sixth leading contributor to attributable disability-adjusted life years (DALYs) in that

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This is an Open Access article distributed under the terms of the Creative Commons Attribution Non-Commercial License (https://creativecommons.org/licenses/ by-nc/4.0) which permits unrestricted non-commercial use, distribution, and reproduction in any medium, provided the original work is properly cited. year.¹ Korea has experienced rapid economic growth in the last century, and the quality of the atmosphere has worsened. Air pollution reduction policies, such as the Special Law on Air Quality in the Seoul metropolitan area, have had limited effect on particulate matter (PM) pollution, and the overall air quality remains poor. Concentrations of PM with aerodynamic diameter less than 10 μ m (PM₁₀) have improved over the past decade, reaching the lowest national average of 45 μ g/m³ in 2012, and then rebounding to a level of 47 μ g/m³ in 2016. However, the concentration of nitrogen dioxide (NO₂) has remained relatively constant, with no large changes. The average values of ozone (O₃) concentration are continuously increasing.²

Epidemiological studies on the health effects of air pollution have been actively conducted in many countries. In particular, time series studies to examine the short-term effects of air pollution have been conducted worldwide and have yielded relatively consistent results.^{3,4} However, cohort studies to assess the long-term effects of air pollution have been primar-

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ily conducted in Western countries that have relatively low concentrations of air pollutants. Due to a lack of direct evidence at higher global concentrations, the integrated expose-response (IER) model was developed. The IER combines information on PM-mortality associations from non-outdoor sources, including secondhand smoke, active smoking, and household air pollution,⁵ and has been used to estimate the disease burden attributable to PM with aerodynamic diameter less than $2.5 \ \mu m (PM_{2.5})$.¹ As the use of IER requires a strict assumption of equal toxicity per unit dose across these non-outdoor sources, cohort studies are needed that reflect the different air pollution concentrations in different regions.⁶

The health effects of air pollution can vary regionally depending on the composition of pollutants or characteristics of the population at risk. The regional differences in PM_{2.5} mortality risk estimates can likely be attributed to geographic variation in particle composition or the spatial heterogeneity of constituents,⁷ as well as differences in the total air pollution mixture.⁸ Regional differences of topography, which may lead to regional differences of exposure error, can contribute to regional differences in PM risk estimates.⁹

To accurately understand the impact of air pollution on health in Korea, the results of research performed specifically for Korea are needed. Since the publication of time series research starting in 1999 in Korea,¹⁰ many epidemiological studies have been conducted; however, the results of these studies have not been systematically summarized. To accurately assess the impact of air pollution in Korea and to clarify future research directions, systematic sorting of epidemiological studies on air pollution conducted in Korea is required. The aim of the present analysis was to comprehensively review previous epidemiological studies on air pollution and health conducted in Korea to identify future study needs.

LITERATURE SEARCH

We conducted a literature search in PubMed using the search terms ("air pollution" [MeSH Terms] OR ("air" [All Fields] AND "pollution" [All Fields]) OR "air pollution" [All Fields]) AND ("mortality" [Subheading] OR "mortality" [All Fields] OR "mortality" [MeSH Terms]) AND ("Korea" [MeSH Terms] OR "Korea" [All Fields]) and (("air pollution" [MeSH Terms] OR ("air" [All Fields] AND "pollution" [All Fields]) OR "air pollution" [All Fields]) AND ("epidemiology" [Subheading] OR "epidemiology" [All Fields] OR "morbidity" [All Fields] OR "morbidity" [MeSH Terms]) AND ("Korea" [MeSH Terms] OR "Korea" [All Fields])) NOT ("mortality" [Subheading] OR "mortality" [MeSH Terms]) to find published studies on the associations of air pollution with mortality and morbidity respectively in Korea, between January 1990 and July 2018.

We also searched for health impact assessment studies using the same search engine and the search terms ("number" [All Fields] AND ("death" [MeSH Terms] OR "death" [All Fields] OR "deaths" [All Fields])) OR "burden of disease" [All Fields] OR "health impact assessment" [All Fields] AND "Korea" [All Fields] AND ("air pollution" [All Fields] OR "ambient" [All Fields]).

After reviewing the title and abstract of each article, we selected epidemiological studies that reported associations between exposure to air pollution and mortality or morbidity. We then summarized these articles according to their characteristics and results.

The initial search for mortality and morbidity returned 87 and 195 results, respectively. After excluding articles that did not meet the inclusion criteria (Fig. 1), there remained 27 (Table 1) and 37 studies (Table 2) on mortality and morbidity, respectively. One of the mortality study also reported morbidity results, so a total of 38 studies were included in the present review. The search for health impact analyses returned 22 studies; four articles remained after a review of titles and abstracts.

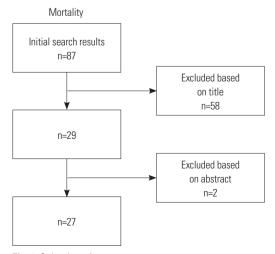
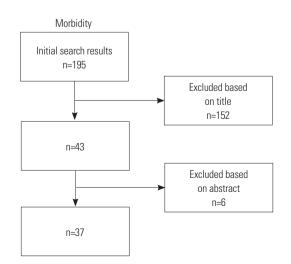


Fig. 1. Selection of papers.



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Author (year)

1 Lee, et al. (1999)¹⁰

2 Hong, et al. (1999)¹³

3 Lee, et al. (1999)¹²

4 Hong, et al. (1999)¹¹

5 Lee, et al. (2000)⁵⁹

6 Kwon, et al. (2001)²⁰

7 Hong, et al. (2002)60

8 Hong, et al. (2002)²¹

9 Ha, et al. (2003)⁶¹

Cohort

No.

| Study design | Study period | Location | Outcome | Pollutant | Unit | Effect size |
|----------------|--------------|----------|-----------------|------------------|-------------------------------|----------------------------|
| | | Seoul | Non-accidental | SO ₂ | 50 ppb | RR 1.078 (1.057, 1.099) |
| | | Ulsan | Non-accidental | SO ₂ | 50 ppb | RR 1.051 (0.991, 1.115) |
| Time series | 1991–1995 | Seoul | Non-accidental | TSP | 100 µg/m³ | RR 1.051 (1.031, 1.072) |
| TITLE SELIES | 1991-1990 | Ulsan | Non-accidental | TSP | 100 µg/m³ | RR 0.999 (0.961, 1.039) |
| | | Seoul | Non-accidental | 1 hr max O_3 | 50 bbp | RR 1.015 (1.005, 1.025) |
| | | Ulsan | Non-accidental | 1 hr max O_3 | 50 bbp | RR 1.020 (0.889, 1.170) |
| Time series | 1995 | Incheon | Total | TSP | 10 µg/m ³ | 1.2% (0.2, 2.2) |
| Time series | 1990 | Incheon | Total | PM ₁₀ | 10 µg/m ³ | 1.2% (0.2, 2.1) |
| | | | Non-accidental | SO ₂ | 50 ppb | RR 1.023 (1.016, 1.084) |
| Case-crossover | | Seoul | Non-accidental | Maximum 03 | 50 ppb | RR 1.023 (0.999, 1.048) |
| | | | Non-accidental | TSP | 100 µg/m³ | RR 1.010 (0.988, 1.032) |
| | | | Total | PM ₁₀ | 10 µg/m ³ | RR 1.007 (1.001, 1.0013) |
| | | | Total | NO ₂ | | RR 1.0026 (1.0006, 1.0046) |
| Time series | 1995–1996 | Incheon | Total | SO ₂ | | RR 1.0023 (0.9996, 1.0051) |
| | | | Total | CO | | RR 1.0019 (0.9990, 1.0049) |
| | | | Total | O ₃ | | RR 0.9951 (0.9908, 0.9994) |
| Time series | 1001 1007 | 7 -:+: | Total | TSP | 100 µg/m³ | 0.5–4% |
| Time series | 1991–1997 | 7 cities | Total | SO_2 | 50 ppb | RR 1.03 (1.01, 1.05) |
| | | | Total | PM ₁₀ | IQR (42.1 µg/m ³) | OR 1.014 (1.006, 1.022) |
| | | | Total | CO | IQR (0.59 ppm) | OR 1.022 (1.017, 1.029) |
| Time series | 1994–1998 | Seoul | Total | NO ₂ | IQR (14.6 ppb) | OR 1.021 (1.014, 1.029) |
| | | | Total | SO ₂ | IQR (9.9 ppb) | OR 1.020 (1.012, 1.028) |
| | | | Total | O ₃ | IQR (20.5 ppb) | OR 1.010 (1.002, 1.017) |
| | | | Stroke | PM ₁₀ | IQR | 1.5% (1.3, 1.8) |
| | | | Stroke | O ₃ | IQR | 2.9% (0.3, 5.5) |
| Time series | 1995–1998 | Seoul | Stroke | NO ₂ | IQR | 3.1% (1.1, 5.1) |
| | | | Stroke | SO_2 | IQR | 2.9% (0.8, 5.0) |
| | | | Stroke | CO | IQR | 4.1% (1.1, 7.2) |
| | | | Ischemic stroke | TSP | IQR | RR 1.03 (1.00, 1.06) |
| | | | Ischemic stroke | SO ₂ | IQR | RR 1.04 (1.01, 1.08) |
| Time series | 1991–1997 | Seoul | Ischemic stroke | NO_2 | IQR | RR 1.04 (1.01, 1.07) |
| | | | Ischemic stroke | CO | IQR | RR 1.06 (1.02, 1.09) |
| | | | | | | |

Table 1. Epidemiological Studies on A

| 10 Kim, et al. (2003) ¹⁵ | Time series | 1995–1999 | Seoul | Non-accidental Respiratory Cardiovascular Cerebrovascular | PM_{10} PM_{10} PM_{10} PM_{10} | IQR (43.12 µg/m ³) IQR (43.12 µg/m ³) IQR (43.12 µg/m ³) IQR (43.12 µg/m ³) | 13.9% (6.8, 21.5) 4.4% (-1.0, 9.0) |
|-------------------------------------|---|-----------|-------|--|--|--|---|
| 11 Kim, et al. (2004) ⁶² | Time series | 1997—2004 | Seoul | Non-accidental Non-accidental | PM ₁₀ (mean) PM ₁₀ (SD) | | RR 1.021 (1.009, 1.035) RR 1.025 (1.000, 1.028) |
| 12 Lee, et al. (2007) ⁶³ | Time series | 2000–2004 | Seoul | Non-accidental | Asian dust event | | Larger effect sizes in the model without Asian dust event |
| 13 Cho, et al. (2008) ⁶⁴ | | 2001 | Seoul | Respiratory | Fine particle count | IQR (10.221 number/cm ³) | 5.73% (5.03, 6.45) |
| 15 GIU, et al. (2008) ²⁴ | et al. (2008) ⁶⁴ Time series | | SEOUI | Respiratory | Respiratory particle count | IQR (10.38 number/cm ³) | 5.82% (5.13, 6.53) |

Ischemic stroke

(postneonates)

Respiratory

Total (postneonates) PM₁₀

03

 PM_{10}

IQR

RR 1.06 (1.02, 1.10)

IQR (42.9 µg/m³) RR 1.142 (1.096, 1.190)

IQR (42.9 µg/m³) RR 2.018 (1.784, 2.283)

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| No. | Author (year) | Study design | Study period | Location | Outcome | Pollutant | Unit | Effect size |
|-----|-----------------------------------|----------------|--------------|----------|---|----------------------------------|---|---|
| | | | | | Infant | PM ₁₀ | 1 μg/m³ | OR 1.000 (0.998, 1.002) |
| | | | | | Infant | NO ₂ | 1 unit | OR 1.002 (0.994, 1.009) |
| 4 | Son, et al. (2008) ²² | Case-crossover | 1999–2003 | Seoul | Infant | SO ₂ | 1 unit | OR 1.015 (0.973, 1.058) |
| | | | | | Infant | CO | 1 unit | OR 1.029 (0.833, 1.271) |
| | | | | | Infant | 0 ₃ | 1 unit | OR 0.984 (0.977, 0.992) |
| | | | | | | | | |
| _ | \/; i (0040) ⁷⁵ | 0 | 0000 0000 | 0 | Non-accidental | PM ₁₀ | 10 μg/m ³ | 0.28% (0.12, 0.44) |
| 5 | Yi, et al. (2010) ⁷⁵ | Case-crossover | 2000-2006 | Seoul | Cardiovascular | PM ₁₀ | 10 µg/m ³ | 0.51% (0.19, 0.83) |
| | | | | | Respiratory | PM ₁₀ | 10 µg/m ³ | 0.59% (-0.08, 1.26) |
| 16 | Kim, et al. (2010) ¹⁹ | Case-crossover | 2004 | 7 cities | Suicide | PM ₁₀ | IQR | 9.0% (2.4, 16.1) |
| 0 | NIIII, et al. (2010) | Case-clossovel | 2004 | 7 cities | Suicide | PM _{2.5} | IQR | 10.1% (2.0, 19.0) |
| 17 | Park, et al. (2011) ⁶⁵ | Time-series | 1999–2007 | Seoul | Non-accidental (high temp. ≥26.2°C) | SO ₂ | 0.5 ppb | 0.83% (0.42, 1.25) |
| | , | | | | Non-accidental | | | |
| | | | | | (low temp. | SO ₂ | 0.5 ppb | 0.21% (0.07, 0.36) |
| | | | | | <26.2°C) | | | |
| | | | | | All-cause infant | TSP | IQR | HR 1.44 (1.06, 1.97) |
| | | | | | All-cause infant | PM ₁₀ | IQR | HR 1.65 (1.18, 2.31) |
| | | | | | All-cause infant | PM _{2.5} | IQR | HR 1.53 (1.22, 1.90) |
| | | | | | All-cause infant | PM _{10-2.5} | IQR | HR 1.19 (0.83, 1.70) |
| 8 | Son, et al. (2011) ⁶⁶ | Birth cohort | 2004–2007 | Seoul | Respiratory infant | TSP | IQR | |
| | | | | | | | | HR 3.78 (1.18, 12.13) |
| | | | | | Respiratory infant | PM ₁₀ | IQR | HR 6.20 (1.50, 25.66) |
| | | | | | Respiratory infant | PM _{2.5} | IQR | HR 3.15 (1.26, 7.85) |
| | | | | | Respiratory infant | PM _{10-2.5} | IQR | HR 2.86 (0.76, 10.85) |
| | | | | | Total | PM ₁₀ | IQR | 0.94% (0.25, 1.62) |
| | | | | | Total | NO ₂ | IQR | 2.27% (1.03, 3.53) |
| | | | | | Total | SO ₂ | IQR | 1.94% (0.80, 3.09) |
| | | | | | Total | CO | IQR | 2.21% (1.00, 3.43) |
| | | | | | | | | |
| 9 | Son, et al. (2012) ¹⁴ | Case-crossover | 2000-2007 | Seoul | Total | 03 | IQR | Positive/NS |
| | | | | | Cardiovascular | PM ₁₀ | IQR | 1.95% (0.64, 3.27) |
| | | | | | Cardiovascular | NO ₂ | IQR | 4.82% (2.18, 7.54) |
| | | | | | Cardiovascular | SO ₂ | IQR | 3.64% (1.46, 5.87) |
| | | | | | Cardiovascular | СО | IQR | 4.32% (1.77, 6.92) |
| | | | | | Cardiovascular | 03 | IQR | Positive/NS |
| | | | | | Non-accidental, | 03 | | 10011100/110 |
| | Heo, et al. (2014) ⁶⁷ | Time-series | 2003–2007 | Seoul | cardiovascular, respiratory | PM _{2.5} and components | | Percentage of excess risk PM _{3.5} and components |
| 1 | Lim, et al. (2014) ⁶⁸ | GWR | 2008–2010 | Seoul | Cardiovascular | PM ₁₀ | | Mean β (SE) 0.956 (0.102 |
| | | | | | Unintentional injury | PM ₁₀ | IQR (48.3 µg/m³) | NS |
| | | | | | Unintentional injury | SO ₂ | IQR (0.005 ppm) | OR 1.119 (1.022, 1.226) |
| 22 | Ha, et al. (2015) ⁶⁹ | Case-crossover | 2002-2008 | 7 cities | Unintentional injury | NO ₂ | IQR (0.02 ppm) | OR 1.208 (1.043, 1.400) |
| | ., | | | | Unintentional injury | 03 | IQR (0.03 ppm) | NS |
| | | | | | | | | |
| 20 | V: 1 1 10047110 | T' · | 1000 0000 | | Unintentional injury | CO | IQR (0.36 ppm) | OR 1.012 (1.000, 1.024) |
| 23 | Kim, et al. (2017) ¹⁶ | Time-series | 1993–2009 | 7 cities | Non-accidental | PM ₁₀ | 10 µg/m ³ | 0.51% (0.01, 1.01) |
| | | | | | Non-accidental | PM_{10} | Daily concentrations of \geq 75 µg/m ³ | 0.48% (0.30, 0.60) |
| 24 | Kim, et al. (2018) ⁷⁰ | Time-series | 1993–2009 | 7 cities | Cardiovascular | PM_{10} | Daily concentrations of \geq 75 µg/m ³ | 0.48% (0.14, 0.82) |
| | | | | | Respiratory | PM_{10} | Daily concentrations of ≥75 µg/m ³ | 1.13% (0.37, 1.89) |

Table 1. Epidemiological Studies on Air Pollution and Mortality in Korea between 1999 and 2018 (Continued)

| No. | Author (year) | Study design | Study period | Location | Outcome | Pollutant | Unit | Effect size |
|-----|----------------------------------|----------------|--------------|-----------|---------------------------------------|-------------------|----------------------|-------------------------------|
| | | | | | Composite cardiovascular events | PM _{2.5} | 1 μg/m³ | HR 1.41 (1.32, 1.50) |
| | | | | | All-cause | PM _{2.5} | 1 μg/m³ | HR 1.32 (1.22, 1.43) |
| | | | | | Cardiovascular | PM _{2.5} | 1 µg/m³ | HR 1.36 (1.11, 1.66) |
| | | | | | Composite cardiovascular events | CO | IQR (0.25 ppm) | HR 1.79 (1.61, 1.99) |
| | | | | | All-cause | CO | IQR (0.25 ppm) | HR 1.72 (1.52, 1.94) |
| | | | | | Cardiovascular | CO | IQR (0.25 ppm) | HR 2.96 (2.12, 4.14) |
| 25 | Kim, et al. (2017) ¹⁸ | Cohort | 2007–2013 | Seoul | Composite cardiovascular events | SO ₂ | IQR (2.54 ppb) | HR 1.94 (1.78, 2.11) |
| | | | | | All-cause | SO ₂ | IQR (2.54 ppb) | HR 1.73 (1.55, 1.92) |
| | | | | | Cardiovascular | SO ₂ | IQR (2.54 ppb) | HR 1.50 (1.14, 1.96) |
| | | | | | Composite cardiovascular events | NO_2 | IQR (18.4 ppb) | HR 2.30 (2.08, 2.55) |
| | | | | | All-cause | NO ₂ | IQR (18.4 ppb) | HR 1.79 (1.59, 2.03) |
| | | | | | Cardiovascular | NO ₂ | IQR (18.4 ppb) | HR 2.67 (1.94, 3.69) |
| | | | | | Composite cardio- vascular events | 03 | IQR (15.9 ppb) | HR 0.63 (0.63, 0.73) |
| | | | | | All-cause | O ₃ | IQR (15.9 ppb) | HR 0.68 (0.63, 0.73) |
| | | | | | Cardiovascular | O ₃ | IQR (15.9 ppb) | HR 0.59 (0.49, 0.71) |
| | | | | | Non-accidental | PM ₁₀ | 10 µg/m ³ | HR 1.05 (0.99, 1.11) |
| | | | | | Cardiovascular | PM ₁₀ | 10 µg/m ³ | HR 1.02 (0.90,1.16) |
| 26 | Kim, et al. (2017) ¹⁷ | Cohort | 2002–2014 | Korea | Cerebrovascular | PM ₁₀ | 10 µg/m ³ | HR 1.14 (0.93, 1.39) |
| 20 | Kiiii, et al. (2017) | CUTUIT | 2002-2014 | KUIEd | Respiratory | PM ₁₀ | 10 µg/m ³ | HR 1.19 (0.91, 1.57) |
| | | | | | Cancer | PM ₁₀ | 10 µg/m ³ | HR 1.02 (0.95, 1.10) |
| | | | | | Lung cancer | PM ₁₀ | 10 µg/m³ | HR 0.96 (0.82,1.13) |
| | | | | | | PM ₁₀ | IQR | Increased OR 1.2% (0.2, 2.3) |
| | | | | | | NO_2 | IQR | Increased OR 4.3% (1.9, 6.7) |
| 27 | Lee, et al. (2018) ⁷¹ | Case-crossover | 2002–2013 | 26 cities | Suicide | SO ₂ | IQR | Increased OR 2.2% (0.7, 3.8) |
| | | | | | | CO | IQR | Increased OR 2.4% (0.9, 3.8) |
| | | | | | | 03 | IQR | Increased OR 1.5% (-0.3, 3.2) |

Table 1. Epidemiological Studies on Air Pollution and Mortality in Korea between 1999 and 2018 (Continued)

TSP, total suspended particles; IQR, interquartile range; OR, odds ratio; RR, relative risk; HR, hazard ratio; NS, not significant; GWR, geographically weighted regression; 7 cities, Seoul, Incheon, Daejeon, Gwangju, Daegu, Busan, Ulsan.

Air pollution and mortality

Among the included studies, the earliest reports regarding an association between air pollution and mortality in Korea were published in 1999.¹⁰⁻¹³ Three of these were time series studies and one was a case-crossover study. Both time series and case-crossover designs are suitable for analysis of acute effects (in days) of short-term exposure to air pollution. One time series analysis was conducted in Seoul and Ulsan. That study reported that the daily variation of ambient concentrations of sulfur dioxide (SO₂), total suspended particles (TSP), and O₃ in Seoul were significantly associated with increased non-accidental mortality.¹⁰ In the same year, the results of reanalysis of Seoul data from the previous time series using a case-crossover ap-

proach, in which each participant became its own control, were reported, showing that only SO₂ was significantly associated with non-accidental mortality.¹² Another time series study conducted in Incheon showed that, in addition to TSP, a 10- μ g/m³ increase in the daily mean concentration of PM₁₀ was also associated with a 1.2% increase in total mortality.¹³ The remaining study was the first to examine the effects of all five criteria pollutants [PM₁₀, SO₂, NO₂, carbon monoxide (CO), and O₃] on mortality in Seoul. That study reported that the previous day's concentrations of PM₁₀ and NO₂ were significantly associated with increased daily mortality [relative risks (RRs) of 1.0007 and 1.0026 for PM₁₀ and NO₂, respectively].¹¹

After 1999, most subsequent studies examined the associa-

tions between air pollutants and total or non-accidental mortality using time series analysis and a case-crossover design. However, the effect sizes varied according to different studies. For instance, the percent increase in mortality for an interquartile range (IQR) increment in PM_{10} ranged between $0.9\%^{14}$ and 3.7%.¹⁵ This may be due to different factors of these studies, including the study period and area, and a multi-city study may provide more robust effect size. There were few multi-city studies and even fewer reported associations with total mortality. The most recent such study stated that a $10-\mu g/m^3$ increase in daily ambient PM_{10} was associated with a 0.51% increase in mortality.¹⁶

The effects of air pollution are not only acute but also chronic, and long-term exposure is generally expected to have a much higher effect size than short-term exposure. However, the chronic effect of air pollution has rarely been examined in Korea. In fact, there were only two studies reporting long-term effects of PM exposure on mortality among our search results, one each for PM₁₀ and PM_{2.5}. Kim, et al.¹⁷ analyzed a sample cohort of the National Health Insurance Service and reported a marginally significant 5% increase in mortality per a 10- μ g/m³ increase in annual PM₁₀ concentration. Another study reported a hazard ratio (HR) of 1.32 for all-cause mortality with an increment of 1 μ g/m³ in PM_{2.5}.¹⁸ Long-term exposure to other gaseous pollutants was also found to be associated with increased risk of mortality, and CO, SO₂, and NO₂ showed HRs of 1.72, 1.73, and 1.79 for each IQR increase, respectively.

The effect of air pollution exposure on mortality is causespecific, and the related cardiovascular and respiratory effects are well known. There have been several reports on cardiovascular and respiratory mortality owing to air pollution in Korea. An interesting cause of death that shows an association with air pollution is suicide. In a case-crossover study conducted using data from seven metropolitan cities in Korea (Seoul, Incheon, Daejeon, Gwangju, Daegu, Busan, and Ulsan), the authors reported that an IQR increase of PM_{2.5} was associated with a 10.1% increase in the number of suicides.¹⁹

Most gaseous air pollutants (SO₂, NO₂, and CO) showed consistently significant associations with increased mortality. For acute exposure, an IQR increase of SO₂, NO₂, and CO increased daily mortality about 2%, and an IQR increase in chronic exposure to those three pollutants showed consistent RRs of around 1.7 (Table 1). However, the association between ambient O₃ concentration and mortality seems inconclusive. Two studies reported significant positive associations of O₃ concentration with total mortality²⁰ and ischemic stroke mortality.²¹ However, we also found reports of significant negative associations with all-cause, ^{13,18} cardiovascular,¹⁸ and infant mortality.²²

Air pollution and morbidity

Asthma and respiratory diseases were among the first specific disorders analyzed in Korea. A time series analysis conducted in Seoul reported that an IQR increase in PM₁₀, SO₂, NO₂, CO,

and O_3 showed significant RRs for children's asthma hospitalization of 1.07, 1.11, 1.15, 1.16 and 1.12, respectively.²³ Another study reported the results of a children's panel for NO₂ exposure showing an OR of 1.12 for upper respiratory symptoms and ORs for lower respiratory symptoms of 1.18, 1.12, and 1.16 for increased exposures to NO₂, SO₂, and CO, respectively.²⁴ O₃ was also associated with children's asthma hospitalization, especially in groups with lower socioeconomic status (RR: 1.32, 95% CI: 1.11, 1.58).²⁵ In a cohort study, O₃ concentration was associated with a 12-month prevalence of wheeze²⁶ and airway hyperresponsiveness²⁷ in children. Other allergic disorders, such as allergic rhinitis and atopic dermatitis, were also associated with air pollution (Table 2).

Similar to the association between air pollution and cardiovascular mortality, the morbidity of cardiovascular and cerebrovascular diseases, such as stroke, myocardial infarction, and hypertension, were also significantly associated with increased exposure to air pollution. A time series analysis reported that NO₂ increased stroke (RR=1.2, *p*-value=0.001),²⁸ and a cohort study reported that long-term exposure to PM_{2.5}, CO, SO₂, and NO₂ increased the risk of acute myocardial infarction, congestive heart failure, and stroke (Table 2).¹⁸

We found two studies examining the association of air pollution with cancer. In these recent studies, indoor radon concentrations were associated with an increased risk of male lung cancer and non-Hodgkin's lymphoma in girls,²⁹ and conventional air pollutants (PM_{10} and NO_2) were associated with lung cancer with marginal significance.³⁰

Similar to the association of suicide with air pollution, depressive symptoms were also found to be associated with air pollution in Korea. A panel study examining air pollution and depressive symptoms was one of the first to report such an association.³¹ An association between PM_{2.5} and major depressive disorder was also found in a community-based urban cohort.³²

Birth outcome has been another subject of analysis. PM_{10} , SO_2 , and CO exposures were reported to have significant associations with low birth weight in a cohort study.^{33,34}

Health impact assessment

Among four studies (Table 3), two calculated the attributable number of deaths,^{35,36} one calculated the attributable number of deaths and morbidity,³⁷ and a fourth calculated DALYs.³⁸

There were substantial differences in the attributable number of deaths among the study results. For instance, Leem, et al.³⁷ estimated the number of deaths attributable to $PM_{2.5}$ to be 15346 in the Seoul metropolitan area, whereas Han, et al.³⁶ estimated this number to be 1763. Yorifuji, et al.³⁵ estimated the number of deaths attributable to PM_{10} over 20 µg/m³ at 5840 in Seoul. These numbers are substantially different, even when considering the differences in study area, study period, and pollutants investigated.

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Table 2. Epidemiological Studies on Air Pollution and Morbidity in Korea between 1999 and 2018

| No. | Author (year) | Study design | Study period | Location | Outcome | Pollutant | Unit | Effect size |
|-----|--|-----------------------|-----------------|----------|--|--|--|--|
| | | | | | Asthma hospitalization | PM ₁₀ | IQR (40.4 µg/m ³) | RR 1.07 (1.04, 1.11) |
| | | | | | Asthma hospitalization | SO_2 | IQR (4.4 ppb) | RR 1.11 (1.06, 1.17) |
| 1 | Lee, et al. (2002) ²³ | Time series | | Seoul | Asthma hospitalization | NO ₂ | IQR (14.6 ppb) | RR 1.15 (1.10, 1.20) |
| | | | | | Asthma hospitalization | 03 | IQR (21.7 ppb) | RR 1.12 (1.07, 1.16) |
| | | | | | Asthma hospitalization | CO | IQR (1.0 ppm) | RR 1.16 (1.10, 1.22) |
| | | | | | Upper respiratory symptoms | NO ₂ | | OR 1.12 (1.01, 1.24) |
| 0 | 1 | | 0000 | 0 | Lower respiratory symptoms | NO ₂ | | OR 1.18 (1.06, 1.31) |
| Ζ | Lee, et al. (2005) ²⁴ | Panel study | 2003 | Seoul | Lower respiratory symptoms | SO ₂ | | OR 1.12 (1.01, 1.25) |
| | | | | | Lower respiratory symptoms | CO | | OR 1.16 (1.02, 1.32) |
| | | | | | Asthma hospitalization (highest SES) | 03 | | RR 1.12 (1.00, 1.25) |
| 3 | Son, et al. (2006) ²⁵ | Time series | 2002 | Seoul | Asthma hospitalization (moderate SES) | 03 | | RR 1.24 (1.08, 1.43) |
| | | | | | Asthma hospitalization (lowest SES) | 03 | | RR 1.32 (1.11, 1.58) |
| 4 | Lee, et al. (2007) ⁷² | Natural experiment | 2002 | Busan | Childhood asthma hospitalization | | RR post Asian game period/ RR baseline | 0.73 (0.49, 1.11) |
| | | | | | Asthma hospitalization | PM ₁₀ | IQR | 31% (14, 51) |
| 5 | Lee, et al. (2006) ⁷³ | Time series | 2002 | Seoul | Asthma hospitalization | SO ₂ | IQR | 29% (8, 53) |
| | | | | | Asthma hospitalization | NO ₂ | IQR | 29% (5, 58) |
| | | | | | Low birth weight | CO | IQR | RR 1.081 (1.002, 1.166 |
| 0 | 0 1 (0007)33 | | 0000 0000 | 0 | Low birth weight | SO ₂ | IQR | RR 1.145 (1.036, 1.267 |
| b | Seo, et al. (2007) ³³ | Cohort | 2002–2003 | Seoul | Low birth weight | PM ₁₀ | IQR | RR 1.053 (1.002, 1.108 |
| | | | | | Low birth weight | NO ₂ | IQR | RR 1.003 (0.954, 1.055 |
| 7 | Moon, et al. (2009) ⁷⁴ | | | 4 cities | Respiratory symptoms | 5 criteria pollutants | | Significant positive association with SO and NO ₂ |
| | | | | Seoul | Low birth weight | PM ₁₀ | | OR 1.08 (0.99, 1.18) |
| | | | | Busan | Low birth weight | PM ₁₀ | | OR 1.24 (1.02, 1.52) |
| | | | | Daegu | Low birth weight | PM ₁₀ | | OR 1.19 (1.04, 1.37) |
| 8 | Seo, et al. (2010) ³⁴ | Cohort | 2004 | Incheon | Low birth weight | PM ₁₀ | | OR 1.12 (0.98, 1.28) |
| | | | | Gwangju | Low birth weight | PM ₁₀ | | OR 1.22 (0.98, 1.52) |
| | | | | Daejeon | Low birth weight | PM ₁₀ | | OR 1.06 (1.00, 1.11) |
| | | | | Ulsan | Low birth weight | PM ₁₀ | | OR 1.19 (1.03, 1.38) |
| | | 2 | | | Cardiovascular hospitalization | PM ₁₀ | 10 µg/m ³ | 0.77% (0.53, 1.01) |
| 9 | Yi, et al. (2010) ^{75*} | Case-crossover | 2001–2006 | | Respiratory hospitalization | PM ₁₀ | 10 µg/m ³ | 1.19% (0.94, 1.44) |
| 10 | Kim, et al. (2011) ²⁶ | Cohort | | | 12-month prevalence of wheeze | | 5 ppb | OR 1.372 (1.016, 1.852 |
| | | | | | Depression (SGDS-K) | PM ₁₀ | IQR | 17.0% (4.9, 30.5) |
| 11 | Lim, et al. (2012) ³¹ | Panel study | | Seoul | Depression (SGDS-K) | NO ₂ | IQR | 32.8% (12.6, 65.6) |
| | | | | | Depression (SGDS-K) | O ₃ | IQR | 43.7% (11.5, 85.2) |
| 12 | Kim, et al. (2012) ⁷⁶ | Panel study | | Seoul | Insulin resistance | PM ₁₀ , O ₃ , NO ₂ | IQR | Significantly increased |
| | | | | | Allergic diseases | Traffic related | Polluted vs. non-polluted | OR 2.12 (1.41, 3.19) |
| 13 | Kim, et al. (2013) ⁷⁷ | Cross sectional | | | | pollutants | school | |
| | Kim, et al. (2013) ⁷⁷ Kim, et al. (2013) ²⁷ | Cross sectional | | | Airway hyperresponsiveness | pollutants O ₃ | | OR 1.60 (1.13, 2.27) |

| No. | Author (year) | Study design | Study period | Location | Outcome | Pollutant | Unit | Effect size |
|-----|---|--------------------------------|-----------------|------------------------|---|--|--|--|
| 15 | Han, et al. (2013) ⁷⁸ | | | | Hemorrhagic fever with renal syndrome | PM ₁₀ | 1 μg/m³ | 0.013 increase of monthly cases |
| | | | | | Allergic disease hospital admission | PM ₁₀ | IQR (30.7 µg/m ³) | 2.2% (0.5, 3.9) |
| | | | | | Asthma hospital admission Respiratory hospital admission | PM ₁₀ PM ₁₀ | IQR (30.7 μg/m ³) IQR (30.7 μg/m ³) | |
| 16 | Son, et al. (2013) ⁷⁹ | | 2003–2008 | 8 cities | Cardiovascular hospital admission | PM_{10} | IQR (30.7 µg/m ³) | 0.7% (0.0, 1.4) |
| 10 | 501, 6t dl. (2015) | | 2003-2000 | 0 01163 | Allergic disease hospital admission | NO_2 | IQR (12.2 ppb) | 2.3% (0.6, 4.0) |
| | | | | | Asthma hospital admission Respiratory hospital admission | NO2 NO2 | IQR (12.2 ppb) IQR (12.2 ppb) | 2.2% (0.3, 4.1) 2.2% (0.6, 3.7) |
| | | | | | Cardiovascular hospital admission | NO_2 | IQR (12.2 ppb) | 2.2% (1.1, 3.4) |
| 17 | Park, et al. (2013) ⁸⁰ | Time series | | 7 cities | Asthma admission | PM ₁₀ , CO, O ₃ , NO ₂ | Children vs. adult | Lower risk in childre for PM ₁₀ and CO |
| 18 | Kim, et al. (2014) ⁸¹ | Cohort | | | Neurodevelopment (MDI) Neurodevelopment (PDI) | PM ₁₀ PM ₁₀ | | β=-2.83; <i>p</i> =0.003 β=-3.00; <i>p</i> =0.002 |
| 19 | Hwang, et al. (2014) ⁸² | Retrospective cohort | | Seoul | Tuberculosis | SO ₂ | IQR | RR 1.07 (1.03, 1.12) |
| 20 | Han, et al. (2015) ²⁸ | Time series | 2004–2013 | | Stroke | NO ₂ | | RR 1.262, <i>p</i> =0.001 |
| 21 | Kim, et al. (2015) ⁸³ | Case-crossover | | Korea | Hourly asthma ED visit | PM _{10-2.5} | IQR | OR 1.05 (1.00, 1.11) |
| 21 | NIII, et al. (2013) | Case-ciossovei | | NUIEd | Hourly asthma ED visit | 03 | IQR | OR 1.10 (1.04, 1.16) |
| 22 | Jang, et al. (2015) ⁸⁴ | Ecological | | Korea | Monthly malaria incidence | NO_2 | | β=-0.884, <i>p</i> <0.01 |
| 23 | Shim, et al. (2016) ⁸⁵ | Cross sectional | 2010–2013 | Korea | Benign prostate hyperplasia | NO ₂ | | OR 2.23 (1.55, 2.39) |
| 24 | Kang at al (2010)86 | Time series | 2006–2013 | Casul | Benign prostate hyperplasia Cardiac arrest | SO ₂ PM _{2.5} | 10 µg/m ³ | OR 2.02 (1.42, 2.88) 1.30% (0.20, 2.41) |
| | Kang, et al. (2016) ⁸⁶ Kim, et al. (2016) ⁸⁷ | Time series Cross sectional | 2000-2013 | 36001 | Allergic rhinitis | CO (during the first year of life) | 100 ppb | OR 1.10 (1.03, 1.19) |
| | | | | | Atopic dermatitis | CO (past 12 months) | 1 ppm | OR 8.11 (1.06, 62.12 |
| | | | | | Asthma | NO_2 | | OR 1.67 (1.03, 2.71) |
| 26 | Kim, et al. (2016) ⁸⁸ | Cross sectional | | | Allergic rhinitis | Black carbon | | OR 1.60 (1.36, 1.90) |
| 20 | | | | | Allergic rhinitis | SO ₂ | | OR 1.09 (1.01, 1.17) |
| | | | | | Allergic rhinitis | NO ₂ | | OR 1.18 (1.07, 1.30) |
| 27 | Han, et al. (2016) ⁸⁹ | Time series | 2004–2014 | Seongdong-gu, Seoul | Intracerebral hemorrhage | PM ₁₀ | | RR 1.09 (1.02, 1.15) |
| 28 | Kim, et al. (2016) ³² | Cohort | 2002–2010 | | Subarachnoid hemorrhage Major depressive disorder | O ₃ PM _{2.5} | 10 µg/m ³ | RR 1.32 (1.10, 1.58) HR 1.44 (1.17-1.78) |
| 20 | Kini, et al. (2010) | ounoit | 2002-2010 | Ruisa | Hypertension | PM ₁₀ | 10 μg/m ³ | OR 1.042 (1.009, 1.0 |
| | | | | | Hypertension in >30 years old | PM ₁₀ | 10 μg/m ³ | OR 1.042 (1.009, 1.0 |
| | | | | | Stroke | PM ₁₀ | 10 μg/m ³ | OR 1.044 (0.979, 1.1 |
| | | | | | Angina | PM ₁₀ | 10 μg/m ³ | OR 0.977 (0.901, 1.0 |
| | | | | | Hypertension | NO ₂ | 10 ppb | OR 1.077 (1.044, 1.1 |
| 29 | Lee, et al. (2016)90 | Cross sectional | 2008-2010 | Korea | Hypertension in >30 years old | NO ₂ | 10 ppb | OR 1.080 (1.043, 1.1 |
| -0 | , | 5.555 55510101 | | | Stroke | NO ₂ | 10 ppb | OR 1.073 (0.994, 1.1 |
| | | | | | Angina | NO ₂ | 10 ppb | OR 1.047 (0.968, 1.1 |
| | | | | | | CO | | |
| | | | | | HVDertension | 60 | | |
| | | | | | Hypertension Hypertension in >30 years old | CO | 10 ppb 10 ppb | OR 1.123 (0.963, 1.3 OR 1.129 (0.963, 1.3 |

Table 2. Epidemiological Studies on Air Pollution and Morbidity in Korea between 1999 and 2018 (Continued)

| Table 2. Epidemiological Studies on Air Pollution and Morbidity in Korea bet | etween 1999 and 2018 (Continued) |
|--|----------------------------------|
|--|----------------------------------|

| No. | Author (year) | Study design | Study period | Location | Outcome | Pollutant | Unit | Effect size |
|-----|------------------------------------|-------------------------------------|-----------------|----------|---|--|--|---|
| 30 | Lee, et al. (2017) ⁹¹ | Cross sectional | 2008–2011 | Korea | Pterygium | PM ₁₀ | 5 µg/m³ | OR 1.23 <i>p</i> =0.023 |
| 21 | Chung, et al. (2017) ⁹² | | | | Cardioembolic stroke | PM_{10} | | Significantly increased |
| 51 | Ghung, et al. (2017) | | | | Cardioembolic stroke | SO ₂ | | Significantly increased |
| 32 | Kim, et al. (2016) ⁹³ | Randomized intervention trial | | | Atopic dermatitis | Indoor VOC | Environmentally friendly vs. PVC wallpaper | More improvement in environmentally friendly wallpaper group |
| | | | | | Male lung cancer | Indoor radon | 10 Bq/m ³ | 1% |
| 33 | Ha, et al. (2017) ²⁹ | Ecological | 1999–2008 | | Female children non-Hodgkin's lymphoma | Indoor radon | 10 Bq/m ³ | 7% |
| 34 | Hwang, et al. (2017) ⁹⁴ | Time series | | | Cardiovascular ED visit | NH4+ (PM _{2.5} component) | | RR 1.05 (1.01, 1.09) |
| | | | | | Acute myocardial infarction | PM _{2.5} | 1 µg/m³ | 1.36 (1.19, 1.56) |
| | | | | | Congestive heart failure | PM _{2.5} | 1 µg/m³ | 1.44 (1.29, 1.61) |
| | | | | | Stroke | PM _{2.5} | 1 µg/m³ | 1.39 (1.27, 1.52) |
| | | | | | Acute myocardial infarction | CO | IQR (0.25 ppm) | 2.12 (1.72, 2.61) |
| | | | | | Congestive heart failure | CO | IQR (0.25 ppm) | 1.86 (1.56, 2.21) |
| | | | | | Stroke | CO | IQR (0.25 ppm) | 2.00 (1.73, 2.30) |
| | | | | | Acute myocardial infarction | SO ₂ | IQR (2.54 ppb) | 1.82 (1.52, 2.19) |
| 35 | Kim, et al. (2017) ^{18*} | Cohort | 2007–2013 | Seoul | Congestive heart failure | SO ₂ | IQR (2.54 ppb) | 2.00 (1.73, 2.32) |
| | | | | | Stroke | SO ₂ | IQR (2.54 ppb) | 2.25 (2.00, 2.54) |
| | | | | | Acute myocardial infarction | NO_2 | IQR (18.4 ppb) | 1.81 (1.46, 2.25) |
| | | | | | Congestive heart failure | NO_2 | IQR (18.4 ppb) | 2.40 (2.02, 2.85) |
| | | | | | Stroke | NO ₂ | IQR (18.4 ppb) | 2.65 (2.29, 3.06) |
| | | | | | Acute myocardial infarction | 03 | IQR (15.9 ppb) | 0.71 (0.63, 0.82) |
| | | | | | Congestive heart failure | 03 | IQR (15.9 ppb) | 0.64 (0.58, 0.71) |
| | | | | | Stroke | 03 | IQR (15.9 ppb) | 0.60 (0.55, 0.65) |
| | Lamichhane, et al. | - · | | | | PM ₁₀ | 10 µg/m ³ | OR 1.09 (0.96, 1.23) |
| 36 | (2017) ³⁰ | Case-control | | Korea | Lung cancer | NO ₂ | 10 ppb | OR 1.10 (1.00, 1.22) |
| | | | | | | Road density | | OR 1.08 (1.01, 1.15) |
| 37 | Yi, et al. (2017) ⁹⁵ | Cross sectional | 2010 | Seoul | Children's atopic eczema | , Road proximity | | OR 1.15 (1.01, 1.31) |
| 38 | Lamichhane, et al. | Dirth ashart | | | Eatal grouth (DDD) | PM_{10} | 10 µg/m³ | -0.26 mm (-0.41, -0.11) |
| აგ | (2018)96 | Birth cohort | | | Fetal growth (BPD) | NO_2 | 10 µg/m ³ | -0.30 mm (-0.59, -0.03) |

SES, socioeconomic status; SGDS-K, Short Geriatric Depression Scale-Korean; MDI, mental developmental index; PDI, psychomotor developmental index; BPD, biparietal diameter; VOC, volatile organic carbon; PVC: polyvinyl chloride; IQR, interquartile range; OR, odds ratio; RR, relative risk; HR, hazard ratio; 7 cities, Seoul, Incheon, Daejeon, Gwangju, Daegu, Busan, Ulsan.

*From the search results of mortality studies.

DISCUSSION

Beginning in 1999, many studies have been conducted to elucidate the health effects of air pollution in Korea. These studies have reported associations with mortality (all-cause, respiratory, cerebrovascular, cardiovascular, infant, injury, and suicide) and morbidity (allergic, respiratory, cardiovascular, cerebrovascular, adverse birth outcomes, depression, and cancer). Most studies examined the short-term effects of air pollution using a time series or case-crossover study design; we found only three cohort studies that examined long-term effects. There were four studies that estimated the health impacts of air pollution, and except for one study that reported DALYs, three studies had inconsistent estimations of the attributable number of deaths.

Estimating health impacts is usually conducted later than other research as previously estimated associations between exposure and outcome, or concentration-response function (C-R function) are required.³⁹ Naturally, the estimated health impact depends on the C-R function used. We suspect that differences in the attributable number of deaths estimated in the three studies reviewed here is partly due to the different

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Table 3. Studies Estimating Health Impact of Air Pollution Conducted in Korea

| No. | Autor (year) | Study period | Location | Outcome | Pollutant | Health impact (person) | | | | | | | | | | |
|-----|---------------------------------------|-----------------|----------------------|--|--|---------------------------|--|--|--|--|--|--|---------|-------------------------------|--|-----|
| | | | | Attributable number of deaths | PM _{2.5} | 15346 | | | | | | | | | | |
| | | | | Attributable number of respiratory hospital admission | PM ₁₀ | 12511 | | | | | | | | | | |
| | | | | Attributable number of cardiovascular hospital admission | PM1 ₀ | 12351 | | | | | | | | | | |
| 1 | $100m \text{ ot al } (2015)^{37}$ | 2010 | Seoul | Attributable number of lung cancer incidence | PM ₁₀ | 1403 | | | | | | | | | | |
| I | Leem, et al. (2015) ³⁷ | 2010 | metropolitan area | Attributable number of asthma attack (children) | PM ₁₀ | 11389 | | | | | | | | | | |
| | | | urou | Attributable number of asthma attack (adults) | PM ₁₀ | 44006 | | | | | | | | | | |
| | | | | Attributable number of chronic bronchitis | PM ₁₀ | 20490 | | | | | | | | | | |
| | | | | Attributable number of acute bronchitis | PM ₁₀ | 278346 | | | | | | | | | | |
| | | | Seoul | Attributable number of deaths | PM ₁₀ over 20 µg/m ³ | 5840 | | | | | | | | | | |
| | | | Busan | Attributable number of deaths | PM ₁₀ over 20 µg/m ³ | 2465 | | | | | | | | | | |
| | | | Daegu | Attributable number of deaths | PM_{10} over 20 μ g/m ³ | 1466 | | | | | | | | | | |
| 2 | Yorifuji, et al. (2015) ³⁵ | 2009 | Incheon | Attributable number of deaths | PM_{10} over 20 μ g/m ³ | 1931 | | | | | | | | | | |
| | | | | | | | | | | | | | Daejeon | Attributable number of deaths | PM_{10} over 20 μ g/m ³ | 599 |
| | | | | | | | | | | | | | | | | |
| | | | Ulsan | Attributable number of deaths | PM_{10} over 20 μ g/m ³ | 539 | | | | | | | | | | |
| 3 | Yoon, et al. (2015) ³⁸ | 2007 | Korea | Disability-adjusted life years | Out door air pollution | 6.89/1000 person | | | | | | | | | | |
| | | | Korea | Attributable number of deaths | PM _{2.5} | 11924 | | | | | | | | | | |
| | | | Seoul | Attributable number of deaths | PM _{2.5} | 1763 | | | | | | | | | | |
| | | | Busan | Attributable number of deaths | PM _{2.5} | 947 | | | | | | | | | | |
| | | | Daegu | Attributable number of deaths | PM _{2.5} | 672 | | | | | | | | | | |
| 4 | Han, et al. (2018) ³⁶ | 2015 | Incheon | Attributable number of deaths | PM _{2.5} | 309 | | | | | | | | | | |
| | | | Gwangju | Attributable number of deaths | PM _{2.5} | 657 | | | | | | | | | | |
| | | | Daejeon | Attributable number of deaths | PM _{2.5} | 342 | | | | | | | | | | |
| | | | Ulsan | Attributable number of deaths | PM _{2.5} | 222 | | | | | | | | | | |
| | | | Sejong | Attributable number of deaths | PM _{2.5} | 49 | | | | | | | | | | |

C-R functions applied by the authors. Specifically, Yorifuji, et al.35 and Leem, et al.37 used C-R functions for mortality derived from epidemiological studies conducted in the United States (U.S.), whereas Han, et al.³⁶ used an IER function developed for the Global Burden of Disease 2010 and 2013. The C-R function derived from U.S. studies only accounted for a relatively low level of PM; thus, it may be inadequate for estimation of health impacts in Korea where exposure to higher concentrations of PM is observed. The IER function was developed by integrating various C-R functions of other exposures, such as tobacco smoke and burning of indoor solid fuel, to fill the gap in exposure range.³⁶ However, it remains uncertain whether the C-R function is comparable to the higher exposure range observed in Korea. Considering this, it is important to produce C-R functions using Korean data to accurately estimate the health impacts of exposure to air pollution.

As mentioned above, the effect of air pollution exposure can be divided into short-term and long-term effects. Typically, short-term effects are examined using time series and casecrossover studies, and long-term effects are investigated in cohort studies. The most recent time series study in Korea reported a 0.51% increase in mortality for each 10- μ g/m³ increase in PM₁₀.¹⁶ This is comparable to the results of a recent meta-analysis of studies from East Asian cities, including Seoul and Incheon, which reported a 0.47% increase in total mortality for the same amount of increase in PM_{10} .⁴⁰ Similarly, although we could not find health impact assessment studies regarding air pollutants other than PM, we believe that previous epidemiological studies can provide relatively robust C-R functions for NO₂ and SO₂ to estimate health impacts.

Previous studies have reported inconsistent associations between O₃ exposure and mortality. Some published studies have reported a negative association, and the cause of this negative association has been an intriguing subject for additional analysis. One hypothesis is that the C-R function between O₃ concentration and mortality is not linear.41 Time series analyses conducted in Korea and Japan support this hypothesis in short-term associations.^{42,43} However, such non-linearity has not been observed in other studies,^{44,45} and the shape of the C-R function between O₃ concentration and acute mortality is still controversial. Nevertheless, studies analyzing the C-R function for long-term exposure of O₃ and mortality consistently report no evidence of a threshold.^{46,47} However, these studies may not have accounted for lower concentrations of O₃; this may be the reason for not observing a non-linear association, as the reported threshold of non-linear associations tends to be at lower concentrations. The negative association reported in a cohort study conducted by Kim, et al.¹⁸ may suggest the existence of a non-linear C-R function between long-term exposure to O_3 and mortality because Korea has lower concentrations of O_3 than the U.S.;⁴³ however, no analysis has been conducted using Korean data, as far as we know.

Among the two cohort studies on air pollution and mortality, one study examined the long-term health effects of PM_{2.5} exposure. Although it is a valuable addition to the current knowledge, the results of that study seem inconsistent with previous reports. For instance, Kim, et al.¹⁸ reported an HR of 1.32 for all-cause mortality for a $1-\mu g/m^3$ increment of PM_{2.5} in a cohort constructed using the National Health Insurance Service database, and a recent U.S. study analyzing a cohort constructed from a Medicare database reported an HR of 1.073 for a 10-µg/m³ increment of PM_{2.5}.⁴⁷ Kim, et al.¹⁸ suggested possible differences in the effect and composition of PM_{2.5}, genetic characteristics, and range of exposure between these studies, although we find a more than 30-fold greater HR difficult to explain. The largest difference between these two studies was in exposure assessment. Kim, et al.¹⁸ linked the concentration measured at a fixed monitoring station to the addresses of participants, whereas Di, et al.47 used a model-based estimation of individual exposure. Another cohort study examined the long-term effect of PM₁₀ exposure.¹⁷ Those authors reported similar effects for PM₁₀ exposure, although the association was not statistically significant. However, this latter study applied an exposure assessment strategy, which could alleviate the effect of misclassification caused by participant mobility and exposure measurement at fixed monitoring stations.

Conventionally, air pollution studies use concentrations measured at fixed monitoring stations for exposure, which is an advantage for providing a large amount of data for a wide range of pollutants. However, data linked to study participants' addresses may not reflect individual exposure, especially when the mobility pattern of individuals is not accounted for.⁴⁸ This limitation may lead to misclassification, which may have substantial implications for the interpretation of results.49 In recent years, advanced sensor and modeling technologies have facilitated individual exposure measurement in air pollution studies with the use of personal sensors and various exposure models based on dispersion models, geographical information, and satellite images.^{48,50,51} Estimation of exposure using these methods in Korea has been reported recently,⁵² and these individual exposure estimation methods should be applied in future studies to reduce uncertainty.

In addition to observational studies, there have been many intervention studies on air pollution and its health effects. Recent intervention studies have explored the benefits of exposure reduction using devices, such as an air purifier^{53,54} and facemasks,⁵⁵ in randomized controlled trials. The strength of intervention studies is two-fold: First, intervention studies may provide more robust evidence regarding the health effects of

exposure to air pollution. Second, these trials may provide evidence regarding the effectiveness of personal measures that can be used to reduce the effects of air pollution. However, due to ethical and practical limitations, randomized controlled trials can only be applied to evaluate acute effects of exposure to air pollution. For instance, it may be unfeasible and unethical to design a study in which a portion of study participants are asked to wear facemasks for a long period (e.g., years). Causal modeling is a method that has been proposed to mitigate the shortcomings of observational studies without the need to conduct a randomized trial. This approach includes marginal structure modeling, instrumental variable analysis, and negative exposure control.⁵⁶ The causal modeling approach provides associations that are free of confounding under certain assumptions, which can be interpreted as causal, similar to the results of a trial. To date, there had been reports on the causal associations of PM2.5, black carbon, and NO2 in various circumstances.^{57,58} Such experimental studies are necessary so as to correctly assess the effects of air pollution on health and to facilitate more effective interventions through which to reduce exposure and to mitigate the health effects of air pollution.

Finally, despite our best efforts to comprehensively summarize the study results regarding the health effects of air pollution exposure in Korea, it is possible that we did not compile a complete list of all relevant research, which should be considered a limitation of the present review.

CONCLUSION

In the present review, we presented epidemiological studies conducted in Korea examining the health effects of exposure to air pollution. For the past 2 decades, there has been a considerable accumulation of knowledge regarding air pollution and health in Korea. However, the present review highlights that additional studies, especially cohort and experimental studies, are needed to provide more robust and accurate evidence that can be used to promote evidence-based policymaking.

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AUTHOR CONTRIBUTIONS

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